



## Anion exchange synthesis of hollow $\beta$ - $\text{In}_2\text{S}_3$ nanoparticles: Adsorption and visible light photocatalytic performances



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### ABSTRACT

Crystal phase and morphology control of nano semiconductor materials play a key role on the optimization of their applications. Here, hollow  $\beta$ - $\text{In}_2\text{S}_3$  nanoparticles (NPs) with higher surface area were synthesized via an anion exchange process under a hydrothermal condition. The adsorption capacity and photocatalytic activity of the hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs were evaluated using Methylene blue (MB) as a model probe both under dark and visible light irradiation. The fast absorption and large adsorption capacity (*ca.* 157.99 mg/g within 60 min) for MB in the solution were associated to the high specific surface area (324.6 m<sup>2</sup>/g) of the  $\beta$ - $\text{In}_2\text{S}_3$  NPs as well as the electrostatic interaction between the  $\beta$ - $\text{In}_2\text{S}_3$  and MB molecule, indicating that adsorption approach plays a crucial role for removing MB from solution. After the adsorption reaction, visible light photocatalytic reaction further degrades the remaining MB in the system. The highest removal efficiency of 100 mg/L of MB solution reached 73.4% under dark for 60 min, and further increased to 92.2% after visible light irradiation for 180 min. The Langmuir isotherm model and pseudo-second-order kinetics were used for the model of the adsorption behavior of MB onto the  $\text{In}_2\text{S}_3$  sample. The as-prepared hollow  $\text{In}_2\text{S}_3$  NPs might be anticipated to be used in dye wastewater treatment.

### 1. Introduction

Water pollution, especially of dye-contaminated water linked to the large-scale development of the textile global industry, has become a substantial matter of concern triggering the research on strategies to remove dyes from contaminated water worldwide [1]. In this context, different nanosized semiconductor materials, including metal oxide and metal sulphide nanomaterials, have been reported as potential active photocatalysts or adsorbents to remove dyes and pigments from contaminated water because of their large specific surface area, unique optical properties, tunable structure and morphology [2,3]. Due to the large energy band gap, metal oxide nanomaterials ( $\text{TiO}_2$ ,  $\text{ZnO}$  etc.) are only active photocatalysts under ultraviolet (UV) light irradiation because the photons in the visible light range have not high enough energy to excite electrons from the valence band to the conduction band [4,5]. Many methods have been used to change the metal oxide electronic structure for visible light absorption, however, these methods also often create structural defects that accelerate the recombination of photo-generated electron-hole pairs, thus limiting the photocatalytic ability and efficiency [6]. As a contrast, chalcogenides such as  $\text{ZnSe}$ ,

$\text{ZnTe}$ , and  $\text{In}_2\text{S}_3$  with narrow energy band gap have shown to be potential active material in light-emitting diodes, optoelectronic devices, solar energy conversion and photocatalysis [7–9]. Up-to-now, various chalcogenides with different morphologies (such as nanotubes, nanorods, nanowires, nanosheets and hollow nanospheres) have been synthesized and used as highly efficient adsorbents/photocatalysts for the removal of organic dyes [10–12]. As a typical III-VI chalcogenide, the stable and non-toxic  $\beta$ - $\text{In}_2\text{S}_3$  (up to 420 °C) with a defect spinel lattice, a relatively narrow band gap of 2.0–2.4 eV and a negative potential (approximate -0.9 eV vs. NHE) has been widely studied in the high-efficient adsorbents/photocatalyst [13–15]. For example, as an n-type of semiconductor,  $\text{In}_2\text{S}_3$  can provide adsorption sites for a strong electrostatic interaction with the cationic dye [16,17]. In the present context, the control of the morphology of nanomaterials has been a key research topic, which plays an important role for improving the performance of different devices [18–20]. In this perspective, owing to their unique physicochemical properties, high density of surface active sites and abundant inner spaces, the interest in the synthesis of hollow nanostructures to be used as delivery vehicles, heterogeneous catalysis and photocatalysis has increased [21]. Recently, reports on hollow  $\beta$ -

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$\text{In}_2\text{S}_3$  spheres announced their good properties for the degradation of dye in solution [22,23]. For instance, Rengaraj et al. [22] reported the high-efficient photodegradation of MB by hollow  $\text{In}_2\text{S}_3$  microspheres prepared using a thiosemicarbazide-assisted solvothermal process. Nayak et al. [23] found that hollow  $\text{In}_2\text{S}_3$  microspheres synthesized by a template-free hydrothermal method exhibited remarkable photocatalytic efficiency for the degradation of MB and crystal violet. On the other hand, Cheng et al. [16] and Sharma et al. [17] reported that  $\beta$ - $\text{In}_2\text{S}_3$  nanospheres also presented considerable adsorption properties for dyes associated to the electrostatic interaction between dyes and the defects at the  $\text{In}_2\text{S}_3$  surface. Furthermore, Natarajan et al. [24] pointed out that a higher adsorption capacity of dye on the porous materials could enhance their photocatalytic efficiency. The combination of the adsorption and photocatalytic properties of hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs for the removal of dyes from aqueous solution is a promising strategy for decontamination. So far, besides the traditional templating, coating and chemical etching methods, the ion exchange process has been proved to be an efficient method to create hollow structure, similar to the Kirkendall effect, which is a diffusion phenomenon at the interface between different solids [25–29]. However, there are only few reports on the control of the morphology of  $\text{In}_2\text{S}_3$  nanostructures by ion exchange method.

Herein, novel hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs were synthesized via ion exchange under the hydrothermal condition without any surfactant or sacrificial agent. Indium chloride ( $\text{InCl}_3$ ) and sodium sulfide ( $\text{Na}_2\text{S}$ ) were employed as indium and sulfur sources, respectively. We proposed a novel formation mechanism of hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs, in which  $\text{InCl}_3$  might be changed to  $\text{In}(\text{OH})_3$  particles in the solution after an ultrasonic treatment at room temperature [30]. Then, due to the different solubility product constant ( $K_{sp}$ ) between  $\text{In}_2\text{S}_3$  and  $\text{In}(\text{OH})_3$ ,  $\text{OH}^-$  anions of  $\text{In}(\text{OH})_3$  NPs could be replaced by  $\text{S}^{2-}$  anions from  $\text{Na}_2\text{S}$  under hydrothermal condition, resulting in the formation of hollow  $\text{In}_2\text{S}_3$  NPs. The detailed structure and optical properties of the synthesized  $\beta$ - $\text{In}_2\text{S}_3$  NPs were studied. The fast adsorption rate and the maximum adsorption capacity (ca. 157.99 mg/g) of  $\beta$ - $\text{In}_2\text{S}_3$  sample for 100 mg/L of MB solution in 60 min were associated to the high specific surface area (324.6  $\text{m}^2/\text{g}$ ) and the strong electrostatic interaction between the n-type  $\text{In}_2\text{S}_3$  and the cationic MB dye, indicating that the adsorption process plays a leading role in the removal of the dye. Moreover, the remaining MB dye in the system after adsorption was further degraded from 73.4% to 92.2% under visible light irradiation for 180 min. Based on the Langmuir isotherm and pseudo-second-order kinetics, the adsorption behavior of MB on  $\text{In}_2\text{S}_3$  can be considered to be a monolayer adsorption process and is related to the electrostatic interaction [31]. On the basis of experimental observations, the hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs obtained from ion exchange process present a high surface area and negatively charged surface [16,17], which is more beneficial for effectively removing the cationic MB and Rhodamine B (RhB) dyes from solution than for removing the anionic Methyl orange (MO) and Congo red (CgR) dyes [32,33].

## 2. Experimental details

### 2.1. Synthesis of hollow $\beta$ - $\text{In}_2\text{S}_3$ nanoparticles

All purchased chemical reagents are analytical grade and used without further purification. The specific synthesis steps of hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs are as follows: 2.00 mmol of  $\text{InCl}_3 \cdot 4\text{H}_2\text{O}$  was dispersed in 20 mL of distilled water at room temperature with continuous magnetic stirring and ultrasonic treatment for 30 min to form the homogeneous suspension of  $\text{In}(\text{OH})_3$  NPs [30]. Then, 20 mL of  $\text{Na}_2\text{S}$  solution containing 5.00 mmol of  $\text{Na}_2\text{S} \cdot 9\text{H}_2\text{O}$  was added dropwise to the  $\text{In}(\text{OH})_3$  suspension. After magnetic stirring for 30 min, a homogeneous pale yellow suspension was obtained. The pale yellow precursor solution was transferred into an autoclave, and heated in a bake oven at 180 °C for 12, 18 and 24 h, respectively. After cooling to room temperature,

the autoclave was opened, and the obtained products were centrifuged and washed several times successively with distilled water and ethanol. The yellow precipitates were dried in a vacuum oven at 70 °C for 6 h and labelled as IS-1, IS-2 and IS-3, corresponding to reaction time of 12, 18 and 24 h, respectively.

### 2.2. Characterization of $\beta$ - $\text{In}_2\text{S}_3$ nanoparticles

X-ray diffraction (XRD) patterns obtained on an X-ray diffractometer (D/Ma x 2500 pc) using  $\text{Cu K}\alpha$  irradiation were used to determine the phase constitution of the prepared samples (Scan rate:  $0.02^\circ \text{ s}^{-1}$ , Accelerating voltage: 40 Kv, and Applied current: 150 mA). Surface composition and electronic states were investigated by energy dispersive spectroscopy (EDS) and X-ray photoelectron spectroscopy (XPS, ESCALAB 250Xi with  $\text{Al-K}\alpha$  source at a power of 180 W (15 kV  $\times$  12 mA)). The morphology of samples was studied by field-emission scanning electron microscope (FESEM, ZEISS MERLIN Compact) and transmission electron microscopy (TEM, Talos F 200 X). The specific BET surface area ( $S_{\text{BET}}$ ) was calculated by following the nitrogen adsorption-desorption and Brunauer-Emmett-Teller method using an automated surface area and pore size analyzer (TriStar II 3020 with analysis bath temperature:  $-195.850^\circ \text{C}$ , warm free space: 11.5981  $\text{cm}^3$  measured, equilibration interval: 5 s, and sample density: 1.000  $\text{g}/\text{cm}^3$ ). UV-vis spectra and diffuse reflectance spectra (DRS) were recorded on a TU-1980 spectrometer equipped with an integrating sphere attachment. In UV-vis DRS,  $\text{BaSO}_4$  was used as the reflectance standard.

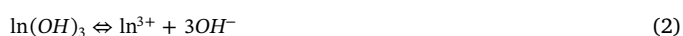
### 2.3. Adsorption and photocatalysis evaluation of $\beta$ - $\text{In}_2\text{S}_3$ nanoparticles

The photocatalytic activity of the  $\text{In}_2\text{S}_3$  sample was evaluated by the degradation of MB solution (100 mg/L) under visible light irradiation (300 W Xenon lamp with a 420 nm cut-off filter). Typically, 40 mg of IS-X (X = 1, 2 or 3) sample was dispersed in 80 ml of MB aqueous solution as the photocatalytic reactor. Then, the reactor was put into a dark room under vigorous stirring for 60 min to ensure the establishment of an adsorption-desorption equilibrium. After adsorption, the mixed suspension was exposed under visible light irradiation for 180 min. At given intervals of time, 3 mL of the mixed suspension was extracted and centrifuged to remove the adsorbent. UV-vis spectroscopy was used to monitor the concentration of the MB solution by measuring the intensity of the MB characteristic absorption at 664 nm. As a contrast, the adsorption property of IS-X (X = 1, 2 or 3) sample for 100 mg/L of MB solution were carried out under dark for 240 min. Meanwhile, the adsorption capacity and the adsorption mechanism were detailed studied in the Supporting Information.

## 3. Results and discussion

### 3.1. Formation mechanism of hollow $\text{In}_2\text{S}_3$ nanoparticles

The formation of the hollow structured  $\text{In}_2\text{S}_3$  NPs can be associated to the different solubility product constant ( $K_{sp}$ ) of  $\text{In}(\text{OH})_3$  ( $1.4 \times 10^{-33}$ ) and  $\text{In}_2\text{S}_3$  ( $6.3 \times 10^{-36}$ ). At the onset of the reaction process, the  $\text{In}(\text{OH})_3$  particles formed by the hydrolysis of  $\text{In}^{3+}$  in aqueous solution can be considered as a template [30]. Due to the large disparity of  $K_{sp}$ , anion exchange following the Kirkendall effect is promoted between  $\text{OH}^-$  and  $\text{S}^{2-}$  anions [28,29], resulting in the formation of  $\text{In}_2\text{S}_3$  hollow nanostructure. The schematic illustration of the formation mechanism of the hollow structured  $\text{In}_2\text{S}_3$  NPs is shown in Fig. 1, and the detail chemical reactions are as follow:



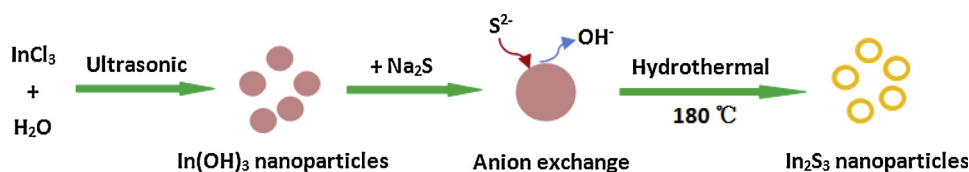


Fig. 1. Schematic illustration of the synthesis and formation mechanism of hollow  $\text{In}_2\text{S}_3$  nanoparticles.



Firstly,  $\text{In(OH)}_3$  particles are formed via the sonication of  $\text{InCl}_3$  aqueous solution at room temperature (Eq. (1)), then there is a weak chemical equilibrium between  $\text{In(OH)}_3$  particles and  $\text{OH}^-$  ions (Eq. (2)). Based on the Kirkendall effect, after adding  $\text{Na}_2\text{S}$  solution into the  $\text{In(OH)}_3$  suspension, mutual diffusion occurs of  $\text{OH}^-$  (outward diffusion) on the surface of  $\text{In(OH)}_3$  particles and  $\text{S}^{2-}$  (inward diffusion) following the Eqs. (2) and (3).  $\text{S}^{2-}$  ions in the solution diffuse inward through the interfacial shell and react with  $\text{In}^{3+}$  ions to form  $\text{In}_2\text{S}_3$  nuclei which precipitated on the surface of  $\text{In(OH)}_3$  NPs. The latter therefore act as a template in the formation process. The  $\text{OH}^-$  anions presenting on the surface of the  $\text{In(OH)}_3$  particles are gradually replaced by  $\text{S}^{2-}$  anions (Eq. (4)), finally forming hollow  $\text{In}_2\text{S}_3$  nanostructures.

### 3.2. Structure and composition

The crystal structure of  $\text{In}_2\text{S}_3$  samples prepared under different reaction time was characterized by XRD as shown in Fig. 2. The observed main diffraction peaks at  $2\theta = 27.4^\circ, 33.3^\circ, 43.7^\circ$  and  $47.8^\circ$  can be indexed to the (311), (400), (511) and (440) planes of the cubic  $\text{In}_2\text{S}_3$  phase (JCPDS card No. 65-0459,  $a = 10.774 \text{ \AA}$ ), similar to previous report [31]. No impurity phases such as  $\text{InS}$ ,  $\text{In}_2\text{O}_3$ , and  $\text{In(OH)}_3$  are observed. Upon increasing the hydrothermal reaction time, the intensity and shape of diffraction peaks, especially (311) plane, sharpen and strengthen, revealing the well crystalline structure and large size of the  $\text{In}_2\text{S}_3$  nanocrystals [34]. Corresponding to the (311) reflection peak, the average grain sizes were calculated using the Scherrer formula to be ca. 5.1 nm, 5.7 nm and 6.9 nm for IS-1, IS-2 and IS-3, respectively [35]. Moreover, the relative intensities of (400) peak to (311) peak ( $I_{400}/I_{311}$ ) and (440) peak to (311) peak ( $I_{440}/I_{311}$ ) of  $\text{In}_2\text{S}_3$  (Table S1) are much higher than that of the standard diffraction pattern, implying that the as-prepared  $\text{In}_2\text{S}_3$  samples grown preferentially along the (400) and (440) directions [18,36].

The typical XPS spectra recorded on the IS-2 sample are shown in Fig. 3. The XPS survey spectrum in Fig. 3A reveals that the sample is mainly composed of In and S element. The traces of C and O observed in the survey spectra might be originated from  $\text{H}_2\text{O}$ ,  $\text{O}_2$  and  $\text{CO}_2$  adsorbed

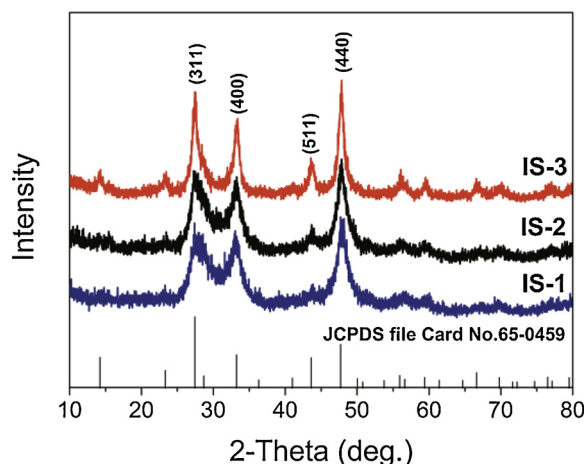


Fig. 2. XRD patterns of the as-prepared  $\text{In}_2\text{S}_3$  samples.

on the surface of the sample due to exposure to the air [37]. In the high energy resolution XPS spectrum recorded on the In element region (Fig. 3B), two high intensity peaks are observed at 445.1 eV and 452.6 eV can be attributed to photoelectrons emitted from the  $\text{In } 3d_{5/2}$  and  $\text{In } 3d_{3/2}$  levels, corresponding to  $\text{In}^{3+}$  oxidation state of  $\text{In}_2\text{S}_3$  [34]. The asymmetry in the In 3d peaks can be fitted by a doublet with components located at 443.8 eV and 451.3 eV, indicating the presence of  $\text{In(OH)}_3$  phase. The absence of  $\text{In(OH)}_3$  in the XRD suggests that the  $\text{In(OH)}_3$  phase is localized on the sample surface, in accordance with the results reported by Zhang et al [38]. The two high-intensity peaks at 161.5 eV and 162.7 eV in the high energy resolution spectrum recorded in the S region (Fig. 3C) correspond to the photoemission from the  $\text{S } 2p_{3/2}$  and  $\text{S } 2p_{1/2}$  levels of  $\text{S}^{2-}$  in the  $\text{In}_2\text{S}_3$  lattice, respectively [23]. The low intensity components at 161.3 eV and 160.1 eV can be associated with an S-rich state for the obtained products, indicating the existence of negatively charged defect points in  $\text{In}_2\text{S}_3$  sample because of the lack of positively charged  $\text{In}^{3+}$  ions in the lattice, resulting in the formation of the negatively charged surface [20].

### 3.3. Morphology

SEM images of the  $\text{In}_2\text{S}_3$  samples obtained under different reaction time are shown in Fig. S1. It can be observed that the  $\text{In}_2\text{S}_3$  samples are composed of aggregated NPs with a narrow size distribution. The average particle sizes of IS-1, IS-2 and IS-3 samples, estimated by Nano measure software [39], gradually increased from 18.7 nm, 23.0 nm to 38.2 nm, suggesting that the hydrothermal reaction time has a significant effect on the size of  $\text{In}_2\text{S}_3$  NPs [40]. Moreover, EDS pattern shown in Fig. S1-D reveals that the sample consists primarily of In and S, and the atomic ratio of In to S is about 1:1.51, in accordance with the XPS results.

Fig. 4 shows typical TEM micrographs, the contrast between the dark edges and lighter central regions in Fig. 4A and B indicates that the  $\text{In}_2\text{S}_3$  NPs actually have a hollow structure. According to the TEM observations, the yield of hollow particles of  $\text{In}_2\text{S}_3$  sample is over 80%. From the high-resolution TEM image (HRTEM) (Fig. 4C) taken at the wall edges of hollow  $\text{In}_2\text{S}_3$  NPs (marked by the white circle in Fig. 3B), parallel lattice fringes can be observed, and the interplanar distance of 0.27 nm corresponded to the (400) lattice plane of cubic  $\beta\text{-In}_2\text{S}_3$  [18]. The selected area electron diffraction (SAED) marked by the white circle is shown in Fig. 4D, the regular circular diffraction rings suggest that the hollow  $\text{In}_2\text{S}_3$  NPs possess polycrystalline structure [10]. Moreover, the (311), (400) and (440) plane are obviously observed in the SAED pattern, consistent with the diffraction peaks in the XRD patterns.

### 3.4. BET measurements

The surface area and pore size distribution of  $\text{In}_2\text{S}_3$  hollow-structured NPs were measured by  $\text{N}_2$  adsorption-desorption isotherm and the BJH method as displayed in Fig. 5. The isotherm patterns can be categorized as a type IV isotherm with H2 hysteresis loop in the range of 0.4–1.0  $P/P_0$ , indicating the presence of mesoporous structure according to the International Union of Pure and Applied Chemistry (IUPAC) classification [41]. The pore size distribution (the inset in Fig. 5) obtained from the desorption branch of the isotherm is broad, and mainly around from 1.5 nm to 15 nm with a unimodal shape at 4.0 nm, which is in agreement with the particle size obtained from TEM



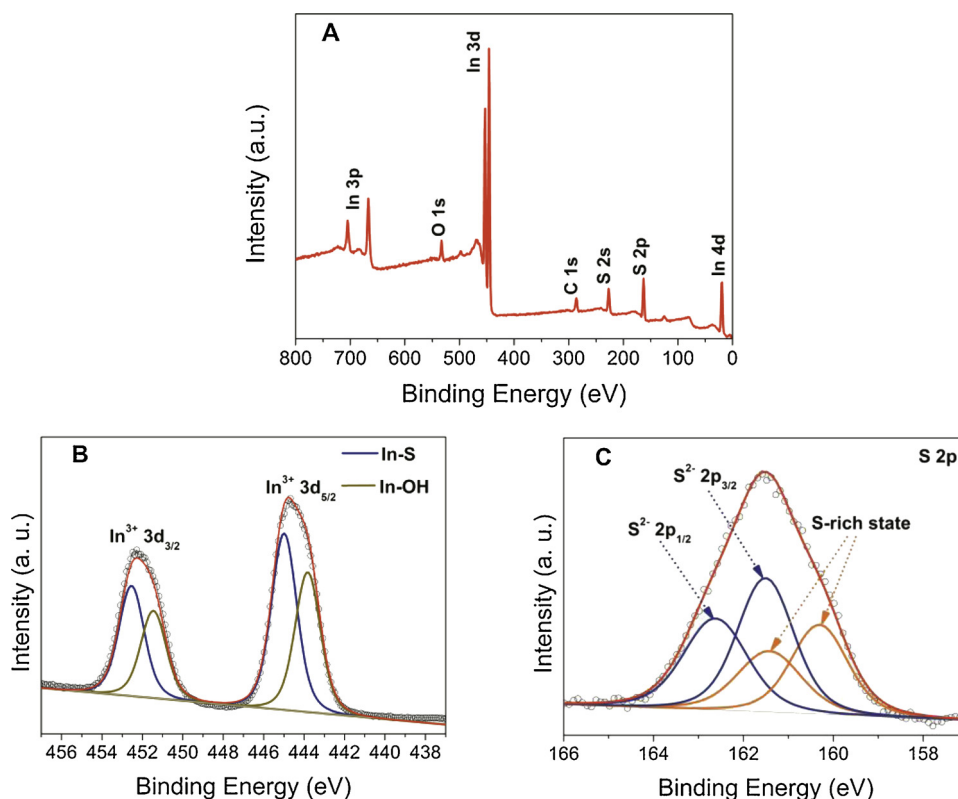


Fig. 3. XPS spectra of sample IS-2: (A) survey, (B) In 3d and (C) S 2p.

analysis. The calculated surface area and average pore diameter of samples IS-1, IS-2 and IS-3 are presented in **Table S2**. The results indicate that the sample IS-2 has the largest surface area of 324.6 m<sup>2</sup>/g, much higher than previously reported In<sub>2</sub>S<sub>3</sub> hollow sphere of 47.4 m<sup>2</sup>/g and 130.7 m<sup>2</sup>/g [22,23]. The large surface area and porous structure of IS-2 are beneficial to remove a large amount of dye molecules in the solution [42].

### 3.5. Optical properties

Fig. 6A shows the UV-vis DRS of the hollow In<sub>2</sub>S<sub>3</sub> NPs in the wavelength range of 200–800 nm. The similar absorption curve with wide and intense absorption peak from 200 to 700 nm indicates that as-prepared In<sub>2</sub>S<sub>3</sub> possesses a strong absorption in the ultraviolet light and visible light regions. Besides, the steep shape of the absorption band

suggests that the visible light absorption of In<sub>2</sub>S<sub>3</sub> NPs is due to the transition of electrons from the valence band to the conduction band, which can be the evidence of the narrow size distribution, uniform crystallites and high phase purity [22].

From these data, the energy band gap can be calculated by extrapolation of the linear portion of the  $(\alpha h\nu)^2$  vs.  $h\nu$  plots to  $\alpha = 0$  (Fig. 6B), using equation:

$$(\alpha h\nu)^2 = A(h\nu - E_g)$$

where  $\alpha$ ,  $h\nu$ , and  $E_g$  are the absorption coefficient, the discrete photon energy, and the band gap energy, respectively. A is a constant [43]. The energy band gaps of samples IS-1, IS-2 and IS-3 were calculated to be ca. 2.15, 2.13 and 2.27 eV, respectively, which is in good agreement with the reported value (2.0–2.3 eV) [23]. These results reveal that the size and surface area of In<sub>2</sub>S<sub>3</sub> NPs resulting from the different reaction

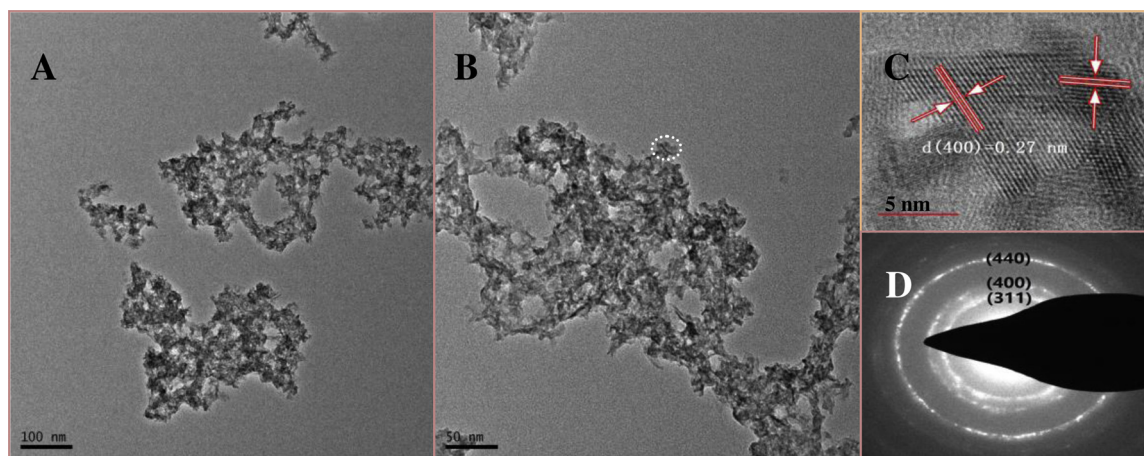


Fig. 4. (A and B) TEM images, (C) high-resolution TEM image, and (D) SAED image of IS-2 sample.

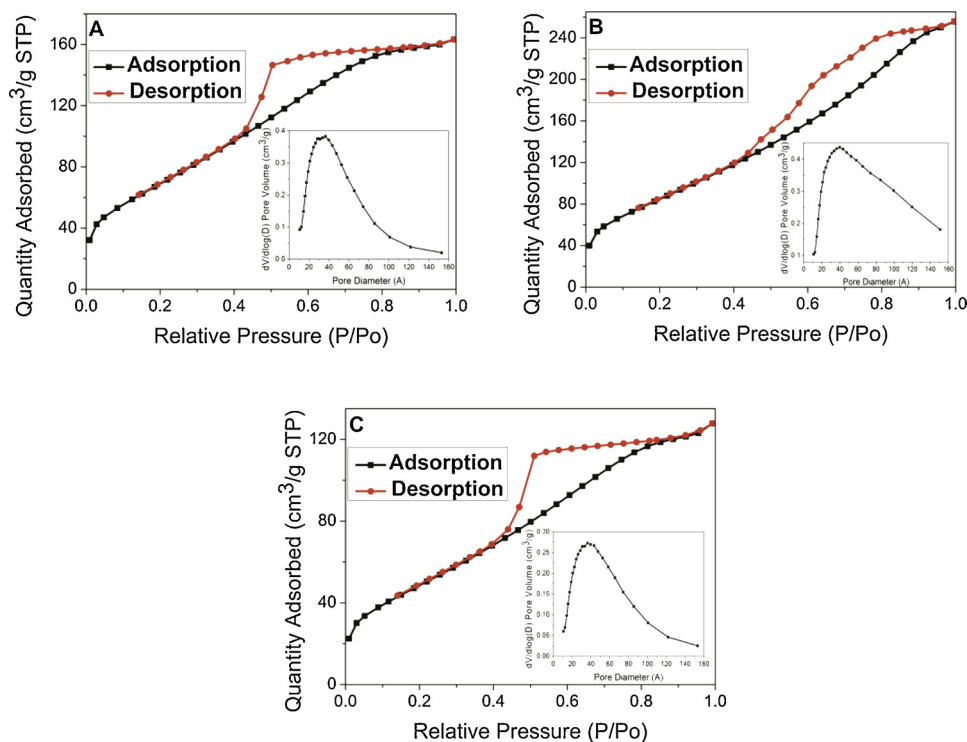


Fig. 5. Nitrogen adsorption-desorption curves: (A) IS-1, (B) IS-2 and (C) IS-3; pore size distribution curves (inset).

time have a significant effect on their optical properties. Meanwhile, IS-2 sample exhibits the strongest visible light absorption, showing a suitable visible light-driven photocatalyst for the degradation of organic dye.

### 3.6. Adsorption and photocatalysis

In the primary adsorption experiments, we found that MB molecules in low concentration solution was nearly completely adsorbed by  $\text{In}_2\text{S}_3$  product in the dark condition, and the solution became clear, indicating no meaning to further identify the photodegradation efficiency. Therefore, the concentrated MB solution (100 mg/L) was used as the model wastewater to be studied in the photocatalytic experiments. Fig. 7A shows the absorbance variation of MB solution as a function of time in the presence of IS-2 sample combining adsorption and photocatalysis. The intensity of the absorbance peak of MB sharply decreased in the dark period suggesting that most of the dye had been adsorbed. After the adsorption, the intensity of absorbance peak of MB further decreased gradually under visible light irradiation, demonstrating the decomposition of the remaining MB dye molecules in the system. From Fig. 7B, the removal% for MB by IS-1, IS-2 and IS-3 after adsorption for

60 min under dark (adsorption-desorption equilibrium) are calculated to be ca. 64.8%, 73.4% and 50.2%, respectively. No more MB dye was adsorbed further increasing the adsorption time to 240 min (Fig. S2). The more removal rate appearing in the adsorption process might be attributed to the strong electrostatic interaction between the dye and catalyst because  $\text{In}_2\text{S}_3$  is an n-type semiconductor material and MB is a cationic dye [16,17,44]. After adsorption-desorption equilibrium, the overall removal of MB onto IS-1, IS-2 and IS-3 under visible light irradiation for 180 min further increased to 82.8%, 92.2% and 63.7%, respectively, indicating that the photocatalytic efficiency under visible light is very significant. Additionally, the photodegradation rate and the corresponding reaction kinetics ( $k$ ) were presented in Table 1. The sample IS-2 exhibits the most efficient photodegradation of 70.8% with the highest  $k$  value of  $0.0060 \text{ min}^{-1}$ , indicating that the surface area and the hollow structure might support more defect sites and active sites, which is beneficial for dark adsorption and visible light photocatalysis of MB in the solution [17,45].

Based on the above experimental analysis, it can be found that the adsorption process plays an important role for removing organic dye. In this context, a series of adsorption experiments of different dyes on IS-2 sample were carried out, in which another cationic RhB dye and two

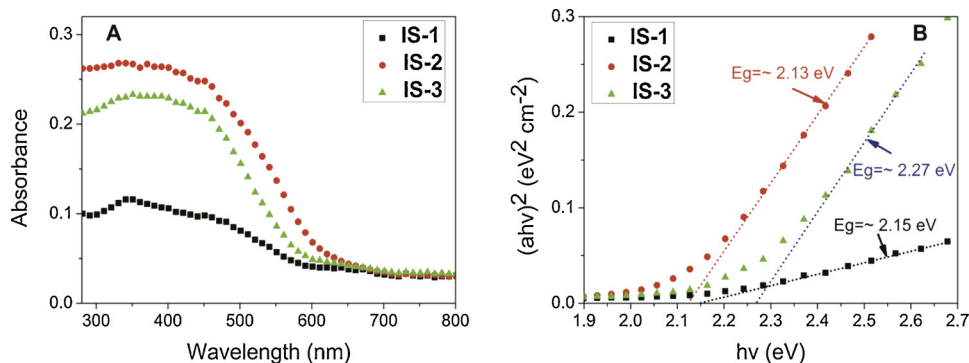


Fig. 6. (A) UV-vis diffuse reflectance spectra and (B) band gaps of  $\text{In}_2\text{S}_3$  hollow nanoparticles.

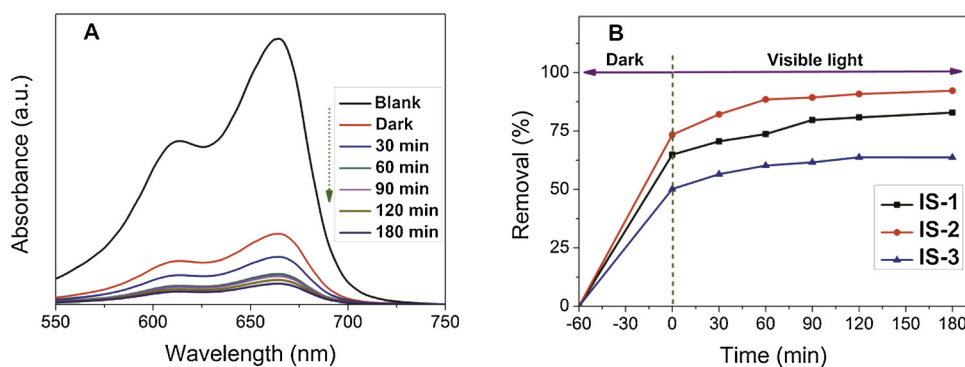


Fig. 7. (A) Absorbance variation of MB solution with time using IS-2 sample as photocatalyst under dark and visible light. (B) Removal% of MB on  $\text{In}_2\text{S}_3$  by adsorption and photocatalytic process.

Table 1

Adsorption capacity, removal/photocatalytic efficiency, and rate constant of 100 mg/L MB solution onto hollow  $\text{In}_2\text{S}_3$  nanoparticles.

Sample	Adsorption capacity (mg/g)	Photocatalytic efficiency (%)	Rate constant ( $k$ ) <sup>a</sup> ( $\text{min}^{-1}$ )	Overall removal efficiency (dark + light) (%)
IS-1	129.6	51.2	0.0041	82.8
IS-2	146.8	70.8	0.0060	92.2
IS-3	100.4	27.1	0.0017	63.7

<sup>a</sup> Calculated by the pseudo-first-order reaction kinetics equation:  $\ln(C/C_0) = -k_1t$  [3].

anionic dyes (MO and CgR) were chosen as the pollutants while the other parameters were kept unchanged. As shown in Fig. S3, the adsorption capacity of IS-2 sample for MO (23.46 mg/g) and CgR (43.83 mg/g) is much lower than that of RhB (100.22 mg/g) and MB (157.99 mg/g), revealing that the electrostatic interaction between adsorbent and dye molecules is favorable for the enhanced adsorption performance [21]. Besides, the adsorption capacity of IS-2 sample for RhB is lower than that of MB dye, which might be associated with the different negative adsorption energy and chromophoric structure of dyes [46]. This result indicates that the hollow  $\text{In}_2\text{S}_3$  NPs can efficiently and selectively adsorb organic dyes, which provides a useful guidance for the photocatalytic degradation of organic dyes of  $\text{In}_2\text{S}_3$  catalyst [47]. Furthermore, in order to accurately evaluate the adsorption capacity and adsorption mechanism of MB onto  $\text{In}_2\text{S}_3$ , IS-2 sample was used as the representative adsorbent to different initial concentration of MB solution (50, 100, 150, 200 mg/L). The corresponding adsorption% and adsorption capacity are shown in Fig. S4 and Table S3. The adsorption speed of the dye is very fast, for the 50 mg/L of MB solution, nearly 100% of MB adsorbed on the particles within 5 min. Upon increasing the dye initial concentration, the adsorption behaviour of the

hollow  $\beta\text{-In}_2\text{S}_3$  for MB repeats the analogous initial fast uptake, almost invariable after 60 min. The calculated maximum adsorption capacity (ca. 157.99 mg/g) is very close to the theoretical value from the pseudo-second-order model and Langmuir isotherm with better correlation coefficient ( $R^2$ ), indicating that the adsorption process of MB onto  $\text{In}_2\text{S}_3$  likely takes place via monolayer molecular adsorption with the rate-limiting step of chemisorptions (Fig. S5-S6 and Table S3-S4).

### 3.7. Photocatalytic mechanism

In order to further understand the photocatalytic mechanism of  $\text{In}_2\text{S}_3$  for MB dye, the valence band (VB) and conduction band (CB) potentials ( $E_{CB}$  and  $E_{VB}$ ) of the hollow  $\beta\text{-In}_2\text{S}_3$  NPs (IS-2) at the point of zero charge have been estimated by the Mulliken equation [46]:

$$E_{VB} = \chi - E_c + 0.5E_g$$

$$E_{CB} = E_{VB} - E_g$$

where  $E_g = 2.13$  eV,  $\chi$  is the electronegativity of the  $\text{In}_2\text{S}_3$  ( $\sim 4.69$  eV),  $E_c$  is the energy of free electrons on the hydrogen scale ( $\sim 4.5$  eV). The calculated values of  $E_{CB}$  and  $E_{VB}$  are ca.  $-0.87$  eV and  $1.26$  eV, respectively. As presented in Fig. 8, the  $E_{VB}$  of  $\text{In}_2\text{S}_3$  sample was more negative than the  $E(\cdot\text{OH}/\text{OH}^-)$  of 2.38 eV and  $E(\cdot\text{OH}/\text{H}_2\text{O})$  of 2.27 eV, suggesting that the photogenerated holes could not convert  $\text{H}_2\text{O}$  to  $\cdot\text{OH}$  and  $\text{H}^+$  [48,49]. Instead, they could directly oxidize the dye molecules into final products [32]. Further, the  $E_{CB}$  of  $\text{In}_2\text{S}_3$  was more negative than  $E(\text{O}_2/\text{O}_2^-)$  of  $-0.28$  eV, so the photogenerated electrons could react with molecular  $\text{O}_2$  absorbed on the surface of  $\text{In}_2\text{S}_3$  photocatalyst to generate  $\cdot\text{O}_2^-$  under visible light irradiation [50]. Finally, the organic dye was converted to small molecules and even  $\text{CO}_2$  and  $\text{H}_2\text{O}$  by  $\cdot\text{O}_2^-$  and  $\text{h}^+$ . Therefore, the  $\cdot\text{O}_2^-$  and  $\text{h}^+$  act as the main active species in the photocatalytic reaction, in agreement with the results of the previous report [51].

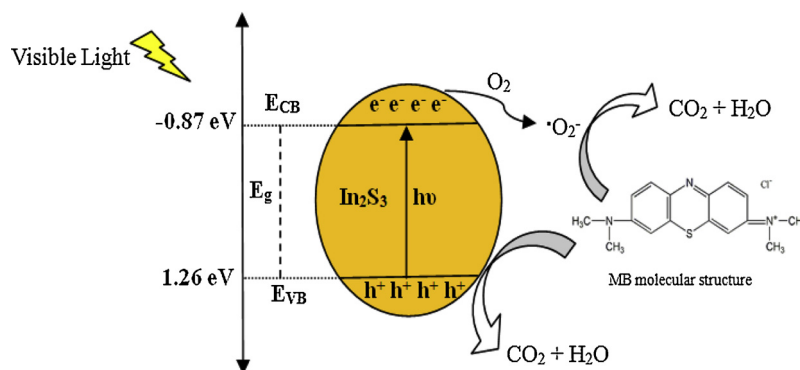


Fig. 8. Schema of the model for the excitation and charge transfer processes of  $\text{In}_2\text{S}_3$  driven by visible light.



#### 4. Conclusions

Hollow-structured  $\beta$ - $\text{In}_2\text{S}_3$  NPs, with a high surface area of  $324.6 \text{ m}^2/\text{g}$ , were synthesized via an anion exchange process, which exhibited highly efficient removal of concentrated MB dye combining dark adsorption and visible light photocatalytic degradation. The highest removal efficiency reached 73.4% by adsorption under dark for 60 min, and further increased to 92.2% after visible light irradiation for 180 min. We observed that the adsorption of dye plays a dominant role in its removal. The enhanced photocatalytic performance observed for the hollow-structured  $\beta$ - $\text{In}_2\text{S}_3$  NPs can be associated to the unique structure and large surface area providing a high density of chemically active sites, higher light-harvesting efficiency, and larger adsorption capacity of the dye. Evidently, hollow  $\beta$ - $\text{In}_2\text{S}_3$  NPs might be exploited for water remediation and many other environmental applications.

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#### Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:<https://doi.org/10.1016/j.jece.2019.102910>.

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